

Preparation and characterization of PVC/Aliquat-336 microspheres for chromium (VI) removal

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ARTICLE INFO

Article History :

Received : 31/03/2019

Accepted : 22/10/2020

Key Words:

PVC; Microspheres;
Cr(VI); Adsorption;
Aliquat-336.

ABSTRACT/RESUME

Abstract: PVC microspheres included with Aliquat-336 as extractant were prepared by phase inversion method. Their morphology and structure was determined by scanning electronic microscopy (SEM) and BET analysis. SEM images show an asymmetric structure and different morphologies, obtained by varying the concentration of the polymer. The prepared microspheres were used as adsorbent to the removal of Cr(VI) from an aqueous solution. Various parameters were studied like, the extractant dosage, the microspheres dosage, the agitation speed and the aqueous solution pH. The prepared microspheres were able to remove chromium from aqueous solution and had a maximum adsorption capacity q_m more than 35mg/g.

I. Introduction

Chromium, a highly toxic element, widely used in several industries must be removed from wastewater before their disposal to natural aquatic environment. For that purpose, several techniques were used for Cr(VI) removal such as reduction-precipitation [1,2,3], solvent extraction [4-6], membrane technologies [7,8,9-13], and electrochemical techniques [14] were described in the literature. However, the adsorption seems to be the most efficient techniques in separation, concentration and removal of chromium. Several adsorbents have been studied for chromium removal from aqueous solutions, including activated carbon [15,16], activated alumina [17], metal oxide [18-20], bimetallic nanocomposites [21], Layered double hydroxide [22], zeolite [23], biomass [24-29], chitosane [30,31], silica gel [32,33], and ions exchanger resins [34-38]. Recently, a new adsorbent constitute of a polymer matrix and an active material, microsphere or microcapsule, gained a great interest in the research of removal of chromium ions [39-50]. Microspheres can be presented as an alternative technology as a new class of adsorbent which can be prepared with a wide range of polymers and various technique.

Several studies were done to extract chromium (VI) from aqueous solutions using microspheres. Polysulfone (PSF) microspheres prepared by phase inversion method using polyvinylpyrrolidone (PVP)

as additive and silicone microspheres prepared by cross-linking in emulsion were studied by N.A. Ochoa et al. [44]. They analyzed their Cr(VI) extraction capacity after their impregnation with Aliquat-336. They found that for a 80 ml solution of 40 ppm the highest extractive performance of Cr(VI) was obtained from 0.3 g microspheres of PSF-PVP (2:1) (98% of extraction at 90 min) and 0.3 g silicone MS (92% of extraction at 90 min). C.O. Illanes et al. [41] studied the kinetics of Cr(VI) adsorption by polysulfone (PSF) and poly(styrene-acrylonitrile) (SAN) microspheres prepared by phase inversion method and impregnated with Aliquat-336. They found Crank model is more realistic and appropriate to describe Cr(VI) kinetic adsorption results. With Crank model, they concluded that there is a surface diffusion contribution to the overall Cr(VI) mass transport inside the microspheres which becomes more important when the microspheres pore size decreases. The removal of Cr(VI) from acidic aqueous solutions using microcapsules with polymeric matrix of ST-EGDMA and a quaternary ammonium salt as the extractant was studied by G. Barassi et al. [40]. The microcapsules were synthesized by adding the extractant Aliquat-336 during the in situ radical copolymerization of the monomers styrene (ST) and ethylene glycol dimethacrylate (EGDMA). They found that, the adsorption kinetic carried out at different temperatures fits well the pseudo-second-order

model with an activation energy of 82.7 kJ mol^{-1} , confirming that the sorption process is controlled by a chemisorption mechanism. The Langmuir's isotherms describe well the adsorption. Finally, from the thermodynamics analysis, they concluded that the chemisorption process is spontaneous and exothermic. S. Ozcan et al. [42] focused their studies on the removal of Cr(VI) from aqueous solution in batch system by using polysulfone microcapsules containing Cyanex 923 as extraction reagent. The microcapsules were prepared by solvent evaporation method. Their results showed that maximum removal of Cr(VI) was obtained by microspheres with Cyanex 923/polysulfone ratio of 1/1. The complex formation between Cr(VI) and Cyanex 923 favoured at initial pH of 1.0. The kinetic adsorption data fits the pseudo second-order kinetic model and the equilibrium time was 30 min. The Redlich Peterson and Langmuir isotherm models describe well the adsorption isotherm data. The Langmuir sorption capacity was $0.430 \text{ mmol.g}^{-1}$. The microspheres preferably adsorbed Cr(VI) against various metal ions, including Cr(III), Ni(II), Pb(II), Cu(II), Zn(II), Cd(II), Co(II). Cellulose microsphere (CMS) adsorbent prepared by radiation-induced grafting of dimethylaminoethyl methacrylate (DMAEMA) onto CMS followed by a protonation process was studied by C. Li et al. [45]. They reported that the adsorption of Cr(VI) onto the resulting adsorbent was very fast; the equilibrium adsorption could be achieved within 15 min. The pH affect strongly the adsorption capacity. A maximum Cr(VI) adsorption (78 mg/g) was obtained at the pH ranging 3.0 to 6.0. The adsorption isotherm data fits the Langmuir isotherm model. Y. Zhang et al. [46] studied the adsorption of Cr(VI) by cellulose microspheres prepared by radiation-induced grafting of styrene onto their surface, followed by acetylation and amination processes. The results showed that adsorption equilibrium could be achieved within 30 min for initial Cr(VI) of 100 mg L^{-1} . The adsorption kinetics was well described by the pseudo-second order model equation, and the adsorption isotherm was better fitted by the Langmuir mode. The maximum theoretical Cr(VI) adsorption capacity was 123.4 mg/g . P. Li et al. [48] prepared polyglycidylmethacrylate (PGMA) microspheres by a dispersion polymerization method and modified by ring-opening reaction to introduce functional hydroxyl groups. Amino groups were grafted to (PGMA) microspheres by cerium initiated surface graft polymerization method and ring-open reaction. They reported that the adsorption of Cr(VI) present a high adsorption capacity (500 mg.g^{-1}). The adsorption isotherm was well fitted with Langmuir isotherm model, and the adsorption process was endothermic and entropy favored in nature. The removal efficiency was optimal at pH1.5. Magnetic poly-(methacrylate-divinyl benzene) microspheres with micron size synthesized by modified

suspension polymerization method and surface modified with amino groups were studied by Wang et al. [49]. The adsorption of Cr(VI) from aqueous solution by the prepared magnetic poly-(MA-DVB) microspheres show that the optimum pH was 3, the contact time was 60 min, and the adsorption capacity increased with increasing temperature. C. Hua et al. [50] studied the adsorption of Cr(VI) by quaternized chitosan microspheres (QCMS) prepared by emulsification solvent extraction method. They concluded that the equilibrium Cr(VI) adsorption capacity achieved within 50 minutes at optimal pH 5. The adsorption kinetics was fitted with pseudo second order model, the adsorption isotherm fitted with Langmuir isotherm model and the adsorption thermodynamic was an endothermic and a spontaneous process. Adsorption of Cr(VI) with mesoporous carbon microspheres prepared by a simple and scalable spray drying method was studied by J. Zhou et al. [47]. They found that at pH 3 maximum adsorption capacity is reached (122 mg/g). Adsorption kinetics for Cr(VI) is well fitted by pseudo-second-order kinetic model. The Langmuir isotherm model gives the best fit to equilibrium experimental data ($q_m=165.3 \text{ mg/g}$). The present work focuses on the preparation and characterization of polymeric microspheres based on PVC, by immersion-precipitation technique, doped with Aliquat-336 and applied for Cr(VI) removal from aqueous solution.

II. Experimental procedures

II.1. Reagents and solutions

Polyvinylchloride (PVC) was purchased from PVC Solvin and used as polymer matrix. Aliquat-336 was purchased from ALDRICH and used as extractant. Dimethylformamide (DMF), supplied by ALDRICH was used as solvent. Sulfuric acid (H_2SO_4), sodium hydroxide (NaOH), potassium chromate (K_2CrO_4) and diphenyl carbazide (DPC) were all analytical grade reagents.

The initial pH of all these aqueous phases was adjusted, with a sulfuric acid solution and measured with a pH meter (InoLab® pH 720).

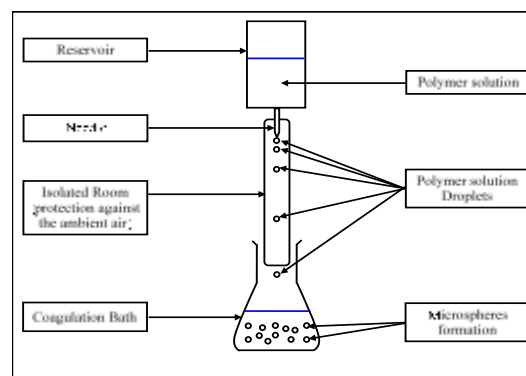


Figure 1. Schematic representation of microspheres setup preparation.

Table 1. polymeric microspheres preparation conditions.

Microsphere	DMF (%w)	PVC (%w)	Ratio Aliquat-336/PVC (w/w)
MS1	94	6	1/1
MS2	90	10	1/1
MS3	86	14	1/1

II.2. Preparation and characterization of microspheres

The microspheres were prepared by immersion-precipitation technique by dissolving poly(vinyl chloride) (PVC) at 6, 10 and 14 % (w/w) in dimethylformamide (DMF). The prepared solutions were maintained under magnetic agitation during 48 hours for a complete dissolution of polymer. Then, quantities of Aliquat-336 equals to that of PVC were added to the solutions previously prepared to have a PVC/Aliquat-336 ratio equal to 1/1. Agitation was continued during 2h. Finally, the solutions were added drop-wise to a bath containing a non-solvent (figure 1). The non-solvent was distilled water at 25°C. Table (1) summarizes the different solutions used to prepare the different microspheres and their attributed names. After the last drop of polymer solution added, the microspheres were collected and transferred to another bath containing distilled water and kept for a day. Then, the microspheres were washed with distilled water and dried in an oven at 40°C temperature overnight.

The microspheres were observed by scanning electron microscopy (SEM) in a FEI QUANTA 200 apparatus, for the section the microsphere was frozen and fractured in nitrogen liquid. The pore diameter, the void fraction, and the surface area were determined by BET surface area analyzer with a Quantachrome NovaWin2 porosimeter.

II.3. Extraction of Cr(VI) by prepared microspheres

The dried microspheres were used for chromium (VI) adsorption from aqueous solutions, in batch system, using a double wall cell at a well known initial concentration, pH and magnetic agitation speed. For chromium analysis, samples of microliters were diluted by distilled water in 25 ml flasks containing 1 ml of diphenylcabazide solution (DPC) in acetone (0.5 g/100ml) and 0.5 ml of

sulfuric acid (5N) and analyzed with UV-Vis spectrophotometer (SCHIMADZU 2101 PC) at 442 nm. Several parameter were investigated such as content of microspheres, initial pH and Cr(VI) concentration and the temperature of the feed solution. The efficiency of adsorption was defined by the following parameters:

The extraction efficiency E (%):

$$E = \left(\frac{C_0 - C_e}{C_0} \right) * 100 \quad (1)$$

The amount of Cr(VI) adsorbed at any time t, q_t (mg/g):

$$q_t = \frac{(C_0 - C_t) * V}{m_1} \quad (2)$$

The amount of Cr(VI) adsorbed at equilibrium q_e (mg/g):

$$q_e = \frac{(C_0 - C_e) * V}{m_1} \quad (3)$$

The microsphere loss of Aliquat-336 L (%):

$$L = \frac{(m_1 - (m_2 - m_{CrO_4} + m_{Cl}))}{m_1} * 100 \quad (4)$$

Where:

m_1 : The microspheres weight before extraction (g);

m_2 : The microspheres weight after extraction (g);

m_{CrO_4} : The weight of chromate adsorbed (g);

m_{Cl} : The weight of chloride desorbed (g);

C_0 : The Cr(VI) initial concentration (mg/l);

C_t : The Cr(VI) concentration at the time t (mg/l).

C_e : The Cr(VI) concentration at equilibrium (mg/l).

III. Results and discussion

III.1. SEM analysis

Figure 2 shows the SEM images of the microspheres prepared by immersion-precipitation technique. Asymmetric spherical microspheres were obtained. The concentration of polymeric solution affects the external morphology of the microspheres. In fact, the roughness of the surface decreases with the increase of polymer solution concentration. The cross section of the microspheres figure (3), show an asymmetrical structure made of a macro-porous internal layer with the presence of macro-void and a thin macro-porous surface layer.

The increase of polymer solution concentration lead to regular macro-pores directed from the surface to the center of microsphere. With the increase of polymer concentration, the precipitation of PVC will occur more easily and the droplets of polymer-poor phase growth by coarsening decrease, so no large macro-void will form.

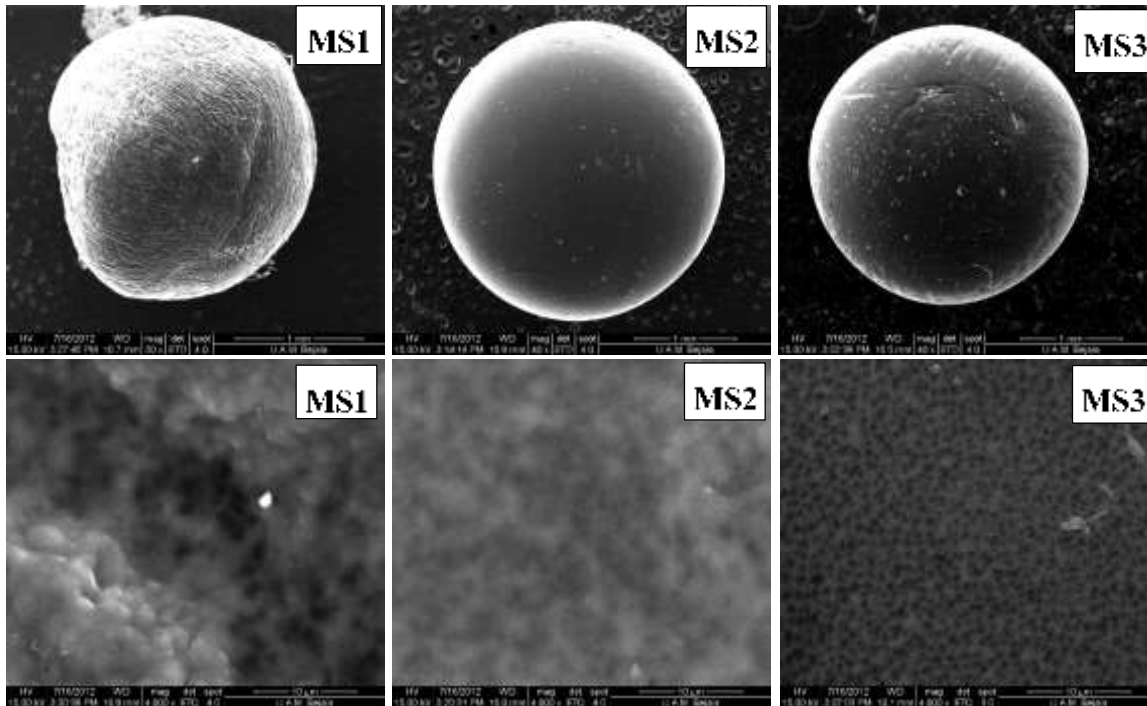


Figure 2. SEM images of different microspheres surfaces and their magnification

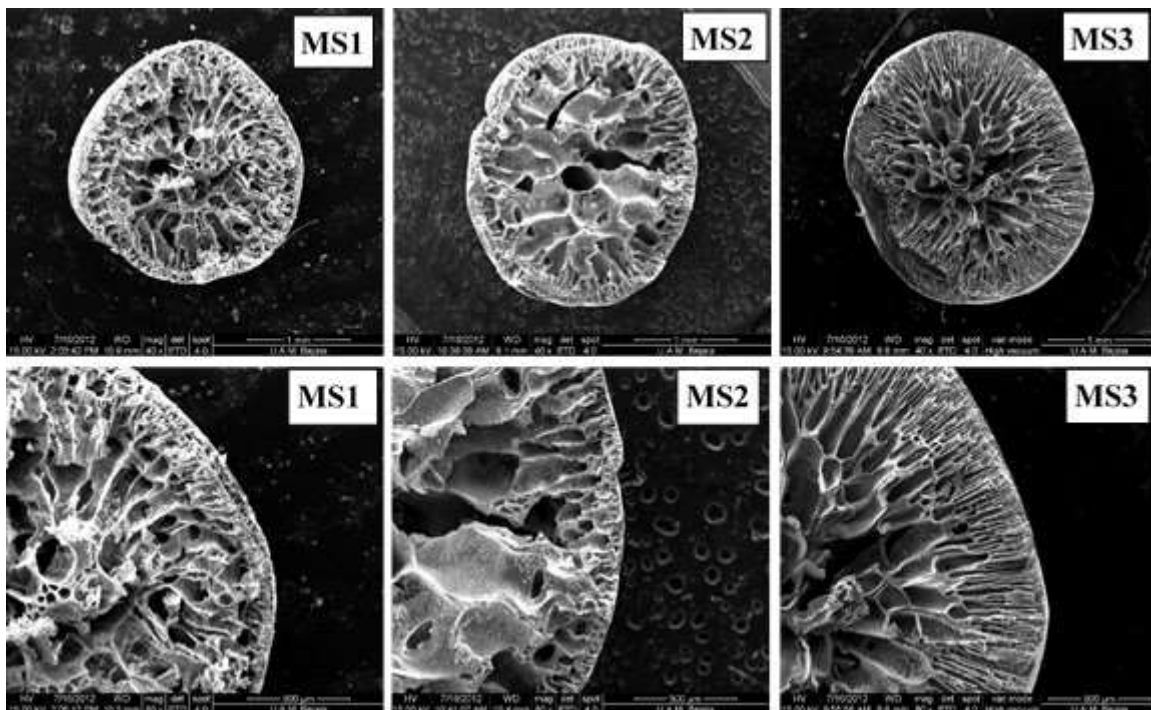


Figure 3. SEM images of the different microspheres cross-sections

III.2. Porosity and diameter of microspheres

The average diameter and apparent density of microspheres were determined by measuring a diameter of known number of microspheres, with a FOJU micrometer and taking the average value, and their weights. The apparent density and porosity of microspheres was calculated, considering a perfect

spherical shape of microspheres, by following equations:

$$\rho_{app} = \frac{m_n}{n \cdot \frac{4}{3} \cdot \pi \cdot r_{avr}^3} \quad (5)$$

$$\varepsilon(\%) = \left(1 - \frac{\rho_{app}}{\rho_p}\right) * 100 \quad (6)$$

Where n , m_n (g), r_{avr} (cm), ρ_{app} (g/cm³), $\varepsilon(\%)$ and ρ_p are the number of microspheres, the weight of the n microspheres, the average diameter of the n

microspheres, the apparent density, the porosity of microspheres and the density of the polymer respectively.

Table 2 summarizes the results of the different measurement. The microspheres diameter have a high value (around 2.5mm) due to the high diameter of the droplets formed at the end of the module and the instantaneous formation of the skin top layer due to the direct contact with the non-solvent. This could be explained by the system used in this study where the droplet in direct contact with air inducing the enlargement of this latter and formation of microsphere of large diameter. The microspheres had a high porosity (great than 70%) because of the presence of micro-void as can be seen in the SEM pictures. The high porosity of microsphere leads to a low apparent density less than 0.32 g/cm³. The increase of polymer concentration decrease microsphere porosity from 82.85 to 73.48 %.

Table 2. *Microspheres average diameter, apparent density and porosity*

	Microspheres average diameter (mm)	Microspheres density (g/cm ³)	Microspheres Porosity (%)
MS 1	2.074	0.2011	82.85
MS 2	2.533	0.2268	80.66
MS 3	2.655	0.3111	73.48

III.3. BET surface area analyses

The microspheres have a small specific surface area and pores volume as shown in table 3. The increase of polymer solution concentration from 6 to 14%, decrease the surface area and the pore volume slightly from 3.561 to 1.857 m²/g and 0.006991 to 0.00362 cm³/g respectively. This could be explained by the excess of Aliquat-336 inducing the decrease in the viscosity of the polymeric solution. As a result, the casting solution cannot retain the spherical shape.

Table 3. *Specific surface area and total pores volume of the different microspheres.*

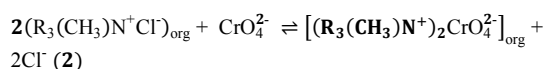
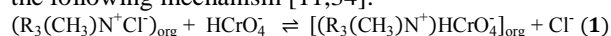
	BET Surface (m ² /g)	BJH Surface (m ² /g)	Total volume of pores (cm ³ /g)
MS1	3.561	2.954	0.006991
MS2	3.431	2.895	0.00594
MS3	1.857	1.028	0.00362

B.J.H: Harett–Joyner–Halenda;

III.4. Adsorption of hexavalent chromium by microspheres

III.4.1. Effect of Aliquat-336 dosage

Figure 4 shows the effect of the Aliquat-336 content in polymeric solution, on the amount of chromium(VI) adsorbed. We notice that for the ratio higher than 1 the droplets of polymeric solution were not formed correctly which can be explained by the excess of Aliquat-336 affecting the formation of microspheres. In fact, the droplet cannot retain the spherical shape induced by a very low viscosity. The adsorption capacity was zero when no extractant was included in the microspheres. The adsorption started with the addition of extractant and increases with the increase of extractant amount added. Thus, no physisorption is produced and adsorption is controlled only by chemisorption with the formation of the Aliquat-336_Chromates complex according to the following mechanism [11,34]:



As the extractant concentration, increased up to about 0.6 g/g Aliquat-336/PVC ratio the adsorption capacity increases slightly, because of the low extractant content in microspheres. Indeed, the extractant diffuses into the coagulation bath during the formation of the microspheres thus a small amount of extractant remain included in the microspheres after their formations. A further increase in the extractant concentration above 0.6 g/g Aliquat-336/PVC ratio, increase exponentially adsorption capacity. This is due to the increases of the extractant content in the microspheres, thus in spite of the diffusion of extractant into the coagulation bath a significant amount of extractant remains included in the microspheres after their formation. For the farther works, an Aliquat-336/PVC ratio of 1 g/g was chosen to prepare the different microspheres.

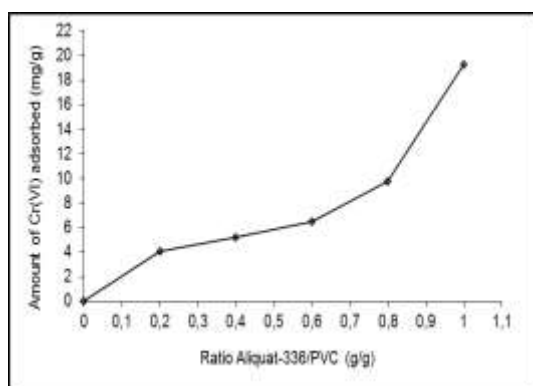


Figure 4. Effect of microspheres MS1 content of extractant on Cr(VI) adsorption (C_0 : 60 mg/l, pH: 7.4, agitation speed: 300 rpm, PVC concentration on polymer solution 6 wt%, contact time: 48h and microspheres dose=1g/L).

III.4.2. Effect of microspheres dosage and polymer concentration

The removal of chromium(VI) by the different microspheres at different adsorbent doses (0.5, 1, 1.5, 2 and 2.5 g/L) for the chromium(VI) concentration of 60 mg/L was studied.

Figure 5-a shows the effect of adsorbent dose on adsorption efficiency of Cr(VI). The figure reveals that adsorption efficiency of Cr(VI) increases rapidly with increasing microspheres dosage from 0.5 to 2.5 g/l, due to adsorption sites with increase in adsorbent dose [34,37,38]. The results in figure 5-b show that the amount of Cr(VI) adsorbed decreased slightly with increase of the microspheres dose. This can be attributed to the fact that some of the adsorption sites remain unsaturated during the adsorption process; whereas the number of available adsorption sites increases by an increase in adsorbent and this results in an increase in removal efficiency [34,37,38]. In the subsequent studies, the microspheres dose will be fixed at 0.5 g/l.

The figure 5 reveals also that MS2, prepared with a solution of 10 wt% of PVC concentration, is more efficient than MS1 and MS3, prepared respectively with a solutions of 6 and 14 wt% of PVC concentration which can be explained by the low porosity and specific surface in the case of microspheres MS3 and to low active sites (extractant content in microspheres) in the case of microspheres MS1 compared to microspheres MS2.

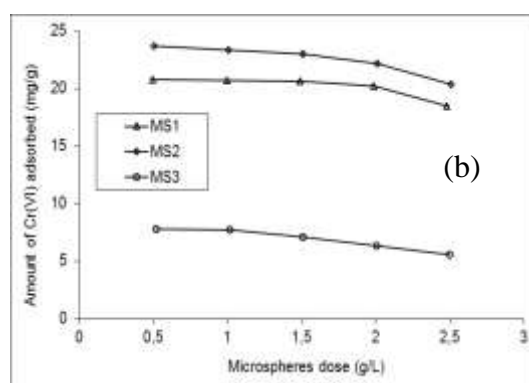
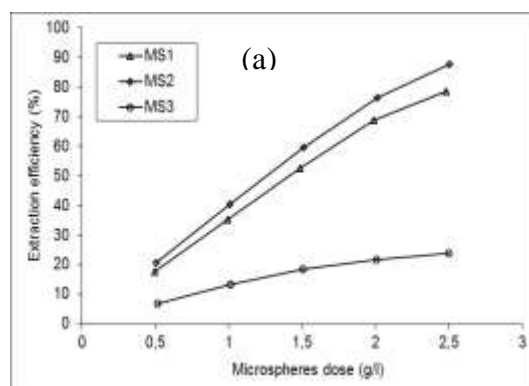


Figure 5. Effect of adsorbent dose on Cr(VI) adsorption efficiency (a) and adsorption amount (b) (C_0 : 60 mg/l, pH: 7.4, agitation speed: 300 rpm and contact time: 48h).

III.4.3. Effect of agitation speed

Agitation speed is an important parameter in sorption phenomena, which has a serious action on the distribution of the solute in the bulk solution and the formation of the external boundary film. The effect of agitation on the uptake of Cr(VI) by the different microspheres was studied at different agitation speeds (200-600 rpm). From figure 6 it can be observed that the increase of agitation speed from 200 to 300 rpm significantly increases the adsorption capacity of Cr(VI) because of best dispersion of microsphere in the aqueous solution and reduction of the diffusion film. The microspheres had a small density below 0.32 kg/l and were hydrophobic so they float on the surface of the aqueous solution at low agitations speeds. Above agitation speed of 300 rpm the adsorption capacity increases slightly because of the increase of the microspheres loss of Aliquat-336 as can be seen on the figure 7. For the further studies, we chose an agitation speed of 300 rpm.

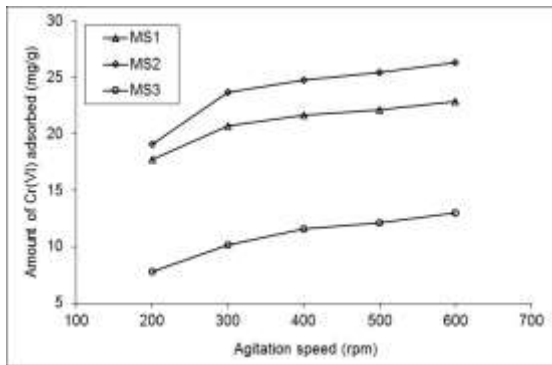


Figure 6. Effect of agitation speed on the adsorption capacity of Cr(VI) by the different microspheres (C_0 : 60 mg/l, pH: 7.4, microspheres dose: 0.5 g/l and contact time: 48h).

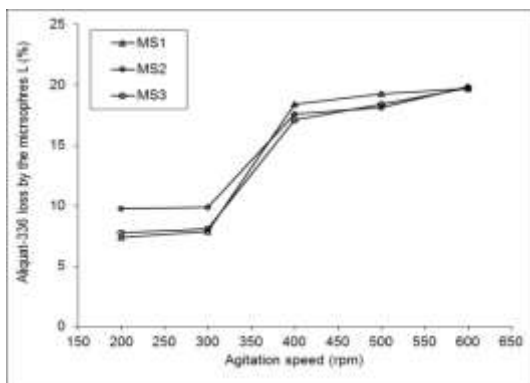
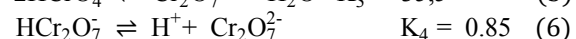
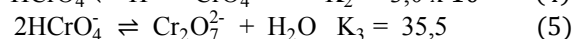
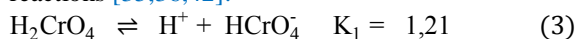


Figure 7. Effect of agitation speed on Aliquat-336 loss by the microspheres (C_0 : 60 mg/l, pH: 7.4, microspheres dose: 0.5 g/l and contact time: 48h).

III.4.4. Effect of pH

Chromium(VI) may exist in the aqueous phase in different forms; the total amount of chromium and the pH are the main variables which govern the chromium species in an aqueous phase [6,52]. The following are the important Cr(VI) equilibrium reactions [35,36,42]:



For the low concentrations, HCrO_4^- is by far the most predominant species for pH between 2 and 5, below pH 2 the amount of HCrO_4^- decrease and H_2CrO_4 rise and above pH 5 the amount of HCrO_4^- decrease and CrO_4^{2-} rise. In case of Aliquat-336 ($\text{R}_3(\text{CH}_3)\text{N}^+\text{Cl}^-$), the extractant reacts as an anion-exchanger forming an ion-pair with the metal species from the aqueous solution, depending on the chromate speciation in solution according to the reactions (1) and (2).

Hence, for a Cr(VI) concentration of 60 mg/l the extraction must reach a maximum at pH between 2 and 5. In our case this not observed; the adsorption capacity of Cr(VI) increases with decreasing the pH (Figure 8). It may be attributed to the microsphere loss of Aliquat-336 as shown in figure 9, were the decrease of pH decreases the amount of Aliquat-336 loss by the microspheres and approach zero value at pH 1. Hence, pH 1 was chosen for the farther studies.

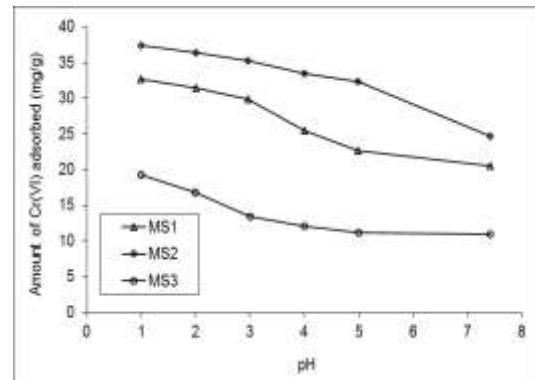


Figure 8. Effect of pH on the adsorption capacity of Cr(VI) by the different microspheres (C_0 : 60 mg/l, agitation speed: 300 rpm, microspheres dose: 0.5 g/l and contact time: 48h).

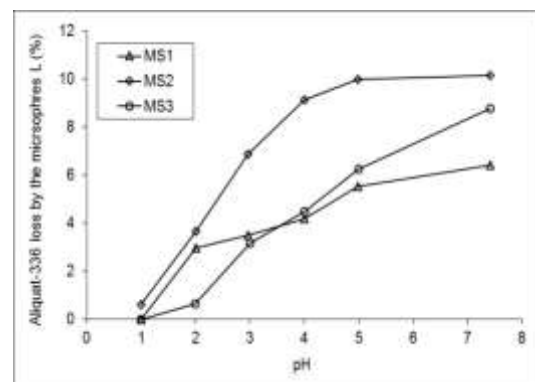


Figure 9. Effect of pH on Aliquat-336 loss by the microspheres (C_0 : 60 mg/l, agitation speed: 300 rpm, microspheres dose: 0.5 g/l and contact time: 48h).

III.4.5. Effect of chromium initial concentration

The initial concentration of Cr(VI) was varied from 5 to 100 mg/L. The uptake of Cr(VI) increased with increase in initial metal concentration (figure 10). At low concentrations, below 40 mg/L, the amount of Cr(VI) adsorbed increase with increasing the chromium initial concentration due to availability of adsorption sites. However, at higher concentration, above 40 mg/L, the amount of Cr(VI) adsorbed reach a plateau due to the saturation of adsorption sites and microspheres reach maximum adsorption capacities.

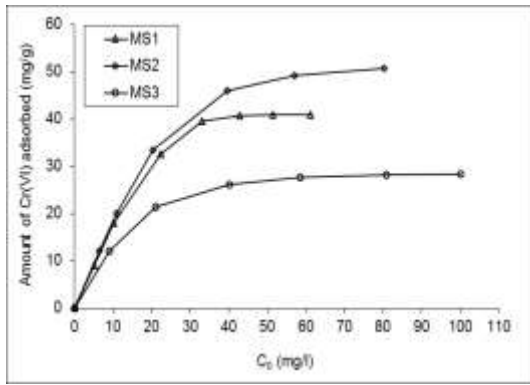


Figure 10. Effect of initial concentration on the adsorption capacity of Cr(VI) by the different microspheres (agitation speed: 300 rpm, microspheres dose: 0.5 g/l, pH =1, contact time: 48h).

III.4.6. Adsorption kinetics

Adsorption processes dependent and controlled by different kinds of mechanisms such as diffusion control, mass transfer, chemical reactions and particle diffusion. The pseudo-first-order kinetic model, the pseudo-second-order kinetic model, the inter-particle diffusion model and the intra-particle diffusion model were used for testing dynamic experimental data at the initial concentrations 10 and 20 mg/l, of Cr(VI) ions. The experimental data was fitted with the theoretical data for each model by minimizing the error Δq (equation (7)) using the solver function in Microsoft Excel. Table (4) summarizes the different kinetics model used to fit the experimental data. The average normalized standard deviation Δq [53-55]:

$$\Delta q(\%) = 100 \sqrt{\frac{\sum [(q_{t,exp} - q_{t,the}) / q_{t,exp}]^2}{(n - 1)}} \quad (7)$$

Where:

$q_{t,exp}$: experimental data of Cr(VI) amount adsorbed at time t;

$q_{t,the}$: theoretical data of Cr(VI) amount adsorbed obtained by the kinetics model at time t.

n: the number of the experimental data.

The model was validated if its average standard deviation $\Delta q(\%)$ is inferior to 5%, the theoretical adsorption capacity at equilibrium $q_{e,the}$ is near to the experimental adsorption capacity at equilibrium $q_{e,exp}$ and in case of diffusion models the constant C must tend to zero. The obtained results were shown in table (5). From the results, the Dumwald-Wagner model seems to be the only model describing the kinetics of adsorption of hexavalent chromium ions by the microspheres. This model has an average normalized standard deviation $\Delta q(\%)$ inferior to 1, the absolute value of the constant C inferior to 10^{-3} and an adsorption capacity at equilibrium near to those of experimental data for all microspheres and for the tow initial concentration of 10 and 20mg/L.

The kinetics adsorption of Cr(VI) by the different microspheres is controlled by intra-particle diffusion through the pores of microspheres and the channels of organic matter (Aliquat-336) inside the polymer matrix. The latter, is occurred by two different mechanisms; the mobile carrier-diffusion mechanism (diffusion of the complex formed (Aliquat-336_Chromate)) or the fixed-site jumping mechanism (complexation reaction between the Aliquat-336-chromate and the adjacent Aliquat-336) [56]. The different constant of the Domwald-Wagner model are shown in table (6) and the different amounts of Cr(VI) adsorbed q_t was depicted in figure 11.

Table 4. The kinetics model equations used to fit the experimental kinetics data.

Model	Kinetic equation	Adsorption capacity $q_{t,the}$
Pseudo-first-order model	$\frac{dq_t}{dt} = k_1(q_e - q_t)$	$q_t = q_e(1 - e^{-k_1t})$
Pseudo-second-order model	$\frac{dq_t}{dt} = k_2(q_e - q_t)^2$	$q_t = \frac{q_e^2 k_2 t}{1 + q_e k_2 t}$
Elovich model	$\frac{dq_t}{dt} = a e^{-bq_t}$	$q_t = \frac{1}{b} (\ln(t) + \ln(ab))$
Intra-particle diffusion : Weber and Morris model	$q_t = k_{id}t^{1/2} + C$	$q_t = k_{id}t^{1/2} + C$
Intra-particle diffusion : Dumwald-Wagner model	$\ln \left[1 - \left(\frac{q_t}{q_e} \right)^2 \right] = -k_{DW}t + c$	$q_t = q_e(1 - e^{-k_1t+c})^{0.5}$

Inter-particle diffusion model	$\ln\left(1 - \frac{q_t}{q_e}\right) = -k_{if}t + c$	$q_t = q_e(1 - e^{-k_1t+c})$
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Table 5. The constants obtained by the fitting of different kinetics models with experimental data for the different microspheres.

C ₀ (mg/L)	Microsphere	Experimental q _e (mg/g)	Kinetics Model						
			Calculated constants	Pseudo-first order	Pseudo-second order	Elovich	Wiber & Moris	Domwal-Wagner	Film diffusion
10	MS1	20.33	q _e (mg/g) Δq (%) C	16.54 18.4377 /	19.06 11.3590 /	/ 8.1741 /	/ 21.7571 3.8235	20.12 0.6278 0.0002	18.99 4.1720 -0.1572
	MS2	22.23	q _e (mg/g) Δq (%) C	17.27 19.6725 /	19.93 12.7311 /	/ 10.2008 /	/ 11.4696 1.6127	21.84 0.7389 -0.0001	20.05 4.6116 -0.1465
	MS3	15.49	q _e (mg/g) Δq (%) C	12.36 17.4233 /	14.26 10.0705 /	/ 6.0438 /	/ 22.9410 0.8229	14.76 0.8727 0.0005	14.15 4.6397 -0.1678
20	MS1	32.65	q _e (mg/g) Δq (%) C	27.14 17.3427 /	30.90 10.6330 /	/ 7.8923 /	/ 14.5000 3.6016	31.79 0.7318 -0.0005	30.28 4.0261 -0.1635
	MS2	34.73	q _e (mg/g) Δq (%) C	28.30 18.7251 /	32.54 12.0055 /	/ 9.6443 /	/ 12.5460 2.9075	34.62 0.5906 -0.0001	32.16 4.9781 -0.1500
	MS3	23.08	q _e (mg/g) Δq (%) C	17.94 16.9064 /	20.38 10.1536 /	/ 7.1955 /	/ 14.8696 2.4616	23.49 0.4161 0.0002	19.91 4.1461 -0.1640

Table 6. The different constants obtained by Dumwald-Wagner models

[Cr(VI)] ₀	Model Constants	Microspheres		
		MS1	MS2	MS3
10 mg/L	k _i (h ⁻¹)	0.0637	0.0507	0.0764
	q _e (mg/g)	20.12	21.84	14.76
	C	0.0002	-0.0001	0.0005
20 mg/L	k _i (h ⁻¹)	0.0699	0.0549	0.0740
	q _e (mg/g)	31.79	34.62	23.49
	C	-0.0005	-0.0001	0.0002

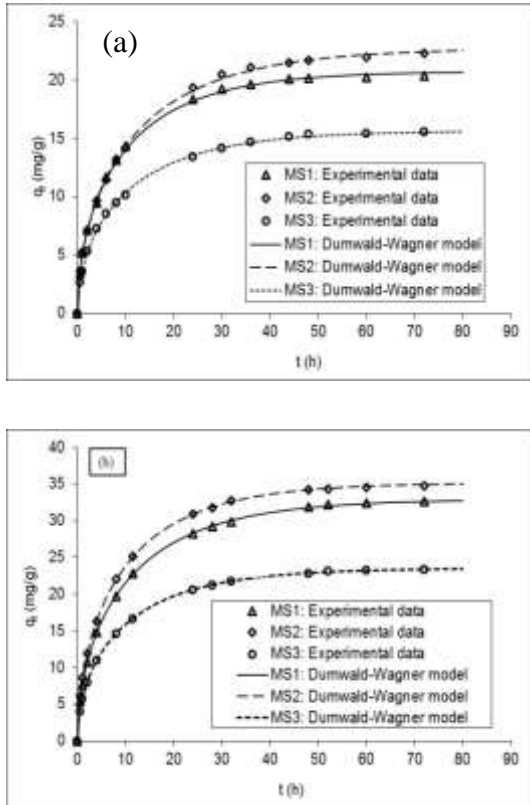


Figure 11. Kinetics of chromium(VI) adsorption by the different microspheres: Symbols are experimental data and lines are the intra-particle diffusion data fitted with Dumwal-Wagner model. Agitation speed: 300 rpm, microspheres dose: 0.5 g/l, pH: 1 and Initial concentration: (a) 10 mg/L and (b) 20mg/L.

In this study, the adsorption of Chromium(VI) on PVC microspheres included with Aliquat-336 is evaluated and compared with popular two and three parameter single solute isotherm models given in table 7. The experimental data was fitted with the theoretical data for each model by minimizing the error Δq (Equation (8)) using the solver function in Microsoft Excel. The average normalized standard deviation Δq [58]:

$$\Delta q(\%) = 100 \sqrt{\frac{\sum[(q_{e-exp} - q_{e-the})/q_{e-exp}]^2}{(n-1)}} \quad (1)$$

Where:

q_{e-exp} : experimental data of adsorption capacity; q_{e-the} : theoretical data obtained from the adsorption isotherm model used; n: the number of the experimental data.

3.4.7. Adsorption isotherm
An adsorption isotherm describes the relationship between the amount of adsorbate that is adsorbed on the adsorbent and the concentration of dissolved adsorbate in the liquid at equilibrium. Several studies with two or three parameter models have been published in the literature to describe experimental data of adsorption isotherms. The Langmuir and Freundlich models are the most frequently used models. The Langmuir model is obtained under the ideal assumption of a totally homogenous adsorption surface, whereas the Freundlich isotherm is suitable for a highly heterogeneous surface. The Redlich–Peterson equation is a combination of Langmuir and Freundlich models [57].

Table 7. The adsorption model equations used to fit the experimental equilibrium data.

Adsorption isotherm Model	Kinetic isotherm equation
Langmuir model	$q_e = \frac{q_m K_L C_e}{1 + K_L C_e}$
Freundlich mode	$q_e = K_F C_e^{1/n}$
Tamkin model	$q_e = \frac{RT}{b_T} \ln(a_T) + \frac{RT}{b_T} \ln(C_e)$
Dubinin–Radushkevich model	$q_e = q_m \exp(-k\varepsilon^2); \quad \varepsilon = RT \ln\left(1 + \frac{1}{C_e}\right)$
Redlich–Peterson model	$q_e = \frac{K_R C_e}{1 + \alpha_L C_e^\beta}$

The model was validated if its average standard deviation $\Delta q(\%)$ is inferior to 5%, the theoretical

maximum amount adsorbed at equilibrium q_{m-the} is near to the experimental maximum amount adsorbed q_{m-exp} . The obtained results were shown in table (8). The different adsorption isotherms obtained experimentally and with Langmuir model are depicted in figure 12.

The adsorption isotherm of chromium (VI) onto microspheres shows a high maximum adsorption capacity more than 35 mg/g (table (8)). Adsorption isotherm models were applied and the results show that Langmuir model describes better the adsorption of chromium(VI) onto the different microspheres as shown in table (8) with an average normalized standard deviation $\Delta q(\%)$ inferior to 1, a maximum adsorption capacity q_m near to that of experimental data and the constant α of Redlich-Peterson isotherm close to 1 indicate that it describes Langmuir model [59,60]. This may be attributed to the chemisorption of Cr(VI) with Aliquat-336; in the way that a chromate reacts with Aliquat-336 so one monolayer adsorption can be obtained and the homogeneity of adsorption sites energy made of Aliquat-366 only with are the assumptions of the Langmuir model [24]. The separation factor R_L , which is defined by the following equation: $1/(1+(1/k_L))$, obtained with Langmuir isotherm ranged between 0.01 to 0.9 for a concentration

between 1 to 100 mg/L indicated that the adsorption is favorable [22,59]. The adsorption energy $E ((-2k_D \cdot R)^{-0.5})$ obtained with Dubinin-Radushkevich model for the different microspheres used (between 13 and 15 kJ/mol) is great then 8 kJ/mol indicate that the adsorption is controlled with chemisorption [22,61,62].

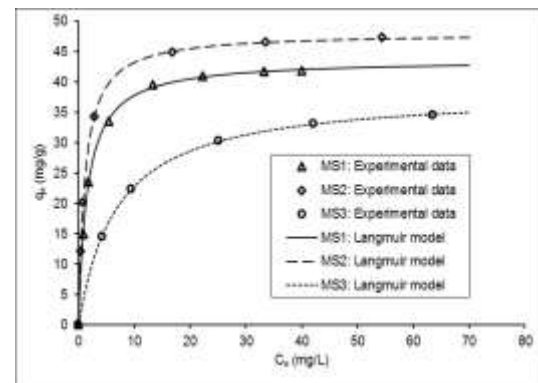


Figure 12. Adsorption isotherm of chromium(VI) onto different microspheres. Agitation speed: 300 rpm, microspheres dose: 0.5 g/l, pH: 1, Temperature: 25°C and contact time: 72h.

Table 8. The constants obtained by the fitting of different isotherm models with experimental data for the different microspheres.

Adsorption isotherm model	Model parameters	Adsorbents		
		MS1	MS2	MS3
Experimental data	q_{m-exp} (mg/g)	41.78	47.30	36.02
Langmuir	k_L (l/mg)	0.6391	0.8933	0.1467
	q_m (mg/g)	43.66	48.03	38.36
	R_L	0.0151-0.1352	0.0111-0.1007	0.0638-0.4053
	Δq (%)	0.6796	0.4956	0.4934
Freundlich	k_F (mg/g (L/mg) ^{1/n})	18.1534	18.7097	10.8817
	n	3.9406	3.7700	3.6039
	Δq (%)	12.2603	18.1597	9.2292
Temkin	a_T (L/g)	10.0923	13.4708	1.9654
	b_T (kJ/mol)	0.3283	0.3124	0.3433
	Δq (%)	7.3033	9.6357	5.3509
Dubinin-Radushkevich	E (kJ/mol)	14.7581	14.6448	13.1077
	k_{D-R} (mol ² /kJ ²)	0.0023	0.0023	0.0029
	q_m (mg/g)	94.27	103.97	77.43
	Δq (%)	10.2181	15.0517	7.4907
Redlich-Peterson	k_R (l/g)	27.9116	43.5114	5.5018
	α_R (l/mg) ^{β}	0.6395	0.9252	0.1372
	β	0.9999	0.9936	1.0102
	Δq (%)	0.6795	0.2726	0.3839

IV. Conclusion

Asymmetric and spherical PVC microspheres included with Aliquat-336 were successfully prepared by immersion-precipitation techniques and efficiently used for chromium (VI) removal from aqueous solution. The morphology and the structure of the prepared microspheres were affected drastically by increasing the polymer concentration from 6 to 14 % in weight.

Chromium(VI) can be effectively removed from aqueous solutions by the prepared microspheres in batch mode. Chromium (VI) uptake by the microspheres was dependent on the pH, agitation speed, microsphere dosage, the microspheres content of Aliquat-336 and polymer concentration. The adsorption capacity was increased with increasing microsphere content of Aliquat-336, agitation speed and hexavalent chromium concentration, but decreased with increasing the pH and the microspheres dosage. The adsorption capacity of Cr(VI) onto microspheres and Aliquat-336 loss also strongly dependent on the solution pH. The decrease of the pH decreases the Aliquat-336 loss by microspheres and no loss observed at pH= 1. The intra-particle model seem to describe the adsorption kinetics of Cr(VI) adsorption by microspheres and Langmuir model describe the thermodynamic of adsorption.

V. References

- Pakzadeh, B.; Batista J.R. Chromium removal from ion-exchange waste brines with calcium polysulfide. *Water Research* 45 (2011) 3055–3064.
- Gheju, M.; Balcu, I. Removal of chromium from Cr(VI) polluted wastewaters by reduction with scrap iron and subsequent precipitation of resulted cations. *Journal of Hazardous Materials* 196 (2011) 131–138.
- Jian Pan, J.; Jiang, J.; Kou Xu, R. Removal of Cr(VI) from aqueous solutions by Na₂SO₃/FeSO₄ combined with peanut straw biochar. *Chemosphere* 101 (2014) 71–76.
- Someda, H.H.; El-Shazly, E.A.; Sheha, R.R. The role of some compounds on extraction of chromium(VI) by amine extractants. *Journal of Hazardous Materials* 117 (2005) 213–219.
- Senol, A. Amine extraction of chromium(VI) from aqueous acidic solutions. *Separation and Purification Technology* 36 (2004) 63–75.
- El-Hefny, N.E. Comparison of liquid-liquid extraction of Cr(VI) from acidic and alkaline solutions by two different amine extractants. *Separation and Purification Technology* 67 (2009) 44–49.
- Bey, S.; Criscuoli, A.; Simone, S.; Figoli, A.; Benamor, M.; Drioli, E. Hydrophilic PEEK-WC hollow fibre membrane contactors for chromium (VI) removal. *Desalination* 283 (2011) 16–24.
- Konczyk, J.; Kozłowski, C.; Walkowiak, W. Removal of chromium(III) from acidic aqueous solution by polymer inclusion membranes with D2EHPA and Aliquat 336. *Desalination* 263 (2010) 211–216.
- Kumbasar, R.A.; Studies on extraction of chromium (VI) from acidic solutions containing various metal ions by emulsion liquid membrane using Alamine 336 as extractant. *Journal of Membrane Science* 325 (2008) 460–466.
- Kozłowski, C.A.; Walkowiak, W. Applicability of liquid membranes in chromium(VI) transport with amines as ion carriers. *Journal of Membrane Science* 266 (2005) 143–150.
- Kebiche-Senhajji, O.; Tingry, S.; Seta, P.; Benamor, M. Selective extraction of Cr(VI) over metallic species by polymer inclusion membrane (PIM) using anion (Aliquat 336) as carrier. *Desalination* 258 (2010) 59–65.
- Basumatary, A.K.; Kumar, R.V.; Ghoshal, A.K.; Pugazhenth, G. Cross flow ultrafiltration of Cr (VI) using MCM-41, MCM-48 and Faujasite (FAU) zeolite-ceramic composite membranes. *Chemosphere* 153 (2016) 436–446.
- Alguacil, F.J.; Lopez-Delgado, A.; Alonso, M. ; Sastre, A.M. The phosphine oxides Cyanex 921 and Cyanex 923 as carriers for facilitated transport of chromium (VI)-chloride aqueous solutions. *Chemosphere* 57 (2004) 813–819.
- Liu, Y.X.; Yuan, D.X.; Yan, J.M.; Li, Q.L.; Ouyang, T. Electrochemical removal of chromium from aqueous solutions using electrodes of stainless steel nets coated with single wall carbon nanotubes. *Journal of Hazardous Materials* 186 (2011) 473–480.
- Barkat, M.; Nibou, D.; Chegrouche, S.; Mellah, A. Kinetics and thermodynamics studies of chromium(VI) ions adsorption onto activated carbon from aqueous solutions. *Chemical Engineerig Process. Process Intensification* 48 (2009) 38–47.
- Adeolu, A.T. ; Okareh, O.T.; Dada, A.O. Adsorption of Chromium Ion from Industrial Effluent Using Activated Carbon Derived from Plantain (Musa paradisiaca) Wastes. *American Journal of Environmental Protection* 4 (2016) 7–20.
- Marzouk, I.; Hannachi, C.; Dammak, L.; Hamrouni, B. Removal of chromium by adsorption on activated alumina. *Desalination and Water Treatment* 26 (2011) 279–286.
- Bhusari, V.N.; Dahake, R.; Rayalu, S.; Bansiwala, A. Comparative Study of Removal of Hexavalent Chromium from Water Using Metal Oxide Nanoparticles. *Advances in Nanoparticles* 5 (2016) 67–74.
- Budiman, F.; Kian, T.W. ; Razak, K.A.; Matsuda, A. ; Lockman, Z. The Assessment of Cr(VI) Removal by Iron Oxide Nanosheets and Nanowires Synthesized by Thermal Oxidation of Iron in Water Vapour. *Procedia Chemistry* 19 (2016) 586–593.
- Nalbandian, M.J.; xZhang, M.J.; Sanchez, J.; Choa, Y.H.; Nam, J. ; Cwiertny, D.M. ; Myung, N. V. Synthesis and optimization of Fe₂O₃ nanofibers for chromate adsorption from contaminated water sources. *Chemosphere* 144 (2016) 975–981.
- Zhou, X.; Jing, G.; Lv, B. ; Zhou, Z. ; Zhu, R. Highly efficient removal of chromium(VI) by Fe/Ni bimetallic nanoparticles in an ultrasound-assisted system. *Chemosphere* 160 (2016) 332–341.
- Khitous, M.; Salem, Z. ; Halliche, D. Effect of interlayer anions on chromium removal using Mg-Al layered double hydroxides: Kinetic, equilibrium and thermodynamic studies. *Chinese Journal of Chemical Engineering* 24 (2016) 433–445.
- Lofù, A. ; Mastroianni, P.; Dell'Anna, M.M. ; Mali, M.; Sisto, R.; Vignola, R. Iron(II) modified natural zeolites for hexavalent chromium removal from contaminated water. *Archives of Environmental Protection* 42 (2016) 35–40.
- Saha, B. ; Orvig, C. Biosorbents for hexavalent chromium elimination from industrial and municipal effluents. *Coordination Chemistry Review* 254 (2010) 2959–2972.
- Jeyaseelan, C. ; Gupta, A. Green Tea Leaves as a Natural Adsorbent for the Removal of Cr (VI) From Aqueous Solutions, Air, Soil. *Water Research* 9 (2016) 13–19.
- Yadav, S.K. ; Ak, D. Efficient Removal of Cr(VI) from Aqueous Solution onto Palm Trunk Charcoal: Kinetic and Equilibrium Studies. *Chemical Science Journal* 7 (2016) 1–7.

27. Henryk, K.; Jaroslaw, C.; Witold, Z. Peat and coconut fiber as biofilters for chromium adsorption from contaminated waste waters. *Environmental Science and Pollution Research* 23 (2016) 527–534.
28. Deepa, K.K.; Sathishkumar, M.; Binupriya, A.R.; Murugesan, G.S.; Swaminathan, K.; xYun, K. Sorption of Cr(VI) from dilute solutions and wastewater by live and pretreated biomass of *Aspergillus flavus*. *Chemosphere* 62 (2006) 833–840.
29. Aksas, H.; Cherifi-Nacy, H.; Babaci, N.; Louahab, K. Etude potentielle de mélange des adsorbants naturels (grignon d'olive et noyaux de dattes) pour l'adsorption du chrome. *Algerian Journal of Environmental Science and Technology* 2 (2016) 48-57.
30. Dima, J.B.; Sequeiros, C.; Zaritzky, N.E. Hexavalent chromium removal in contaminated water using reticulated chitosan micro/nanoparticles from seafood processing wastes. *Chemosphere* 141 (2015) 100–111.
31. Li, K.; Li, P.; Cai, J.; Xiao, S.; Yang, H.; Li, A. Efficient adsorption of both methyl orange and chromium from their aqueous mixtures using a quaternary ammonium salt modified chitosan magnetic composite adsorbent. *Chemosphere*. 154 (2016) 310–318.
32. Liu, Y.; Guo, L.; Zhu, L.; Sun, X.; Chen, J. Removal of Cr(III, VI) by quaternary ammonium and quaternary phosphonium ionic liquids functionalized silica materials. *Chemical Engineering Journal* 158 (2010) 108–114.
33. Pérez-Quintanilla, D.; Del Hierro, I.; Fajardo, M.; Sierra, I. Cr(VI) adsorption on functionalized amorphous and mesoporous silica from aqueous and non-aqueous media. *Materials Research Bulletin* 42 (2007) 1518–1530.
34. Pehlivan, E.; Cetin, S. Sorption of Cr(VI) ions on two Lewatit-anion exchange resins and their quantitative determination using UV-visible spectrophotometer. *Journal of Hazardous Materials* 163 (2009) 448–453.
35. Kabay, N.; Arda, M.; Saha, B.; Streat, M. Removal of Cr(VI) by solvent impregnated resins (SIR) containing aliquat 336. *Reactive and Functional Polymers* 54 (2003) 103–115.
36. Saha, B.; Gill, R.J.; Bailey, D.G.; Kabay, N.; Arda, M. Sorption of Cr(VI) from aqueous solution by Amberlite XAD-7 resin impregnated with Aliquat 336. *Reactive and Functional Polymers* 60 (2004) 223–244.
37. Shi, T.; Wang, Z.; Liu, Y.; Jia, S.; Changming, D. Removal of hexavalent chromium from aqueous solutions by D301, D314 and D354 anion-exchange resins. *Journal of Hazardous Materials* 161 (2009) 900–906.
38. Rengaraj, S.; Joo, C.K.; Kim, Y.; Yi, J. Kinetics of removal of chromium from water and electronic process wastewater by ion exchange resins: 1200H, 1500H and IRN97H. *Journal of Hazardous Materials* 102 (2003) 257–275.
39. Yang, W.W.; Luo, G.S.; Gong, X.C. Polystyrene microcapsules containing Aliquat 336 as a novel packing material for separation of metal ions. *Hydrometallurgy*. 80 (2005) 179–185.
40. Barassi, G.; Valdés, A.; Araneda, C.; Basualto, C.; Sapag, J.; Tapia, C.; Valenzuela, F. Cr(VI) sorption behavior from aqueous solutions onto polymeric microcapsules containing a long-chain quaternary ammonium salt: Kinetics and thermodynamics analysis. *Journal of Hazardous Materials* 172 (2009) 262–268.
41. Illanes, C.O.; Ochoa, N.A.; Marchese, J. Kinetic sorption of Cr(VI) into solvent impregnated porous microspheres. *Chemical Engineering Journal* 136 (2008) 92–98.
42. Ozcan, S.; Tor, A. M.E. Aydin, Removal of Cr(VI) from aqueous solution by polysulfone microcapsules containing Cyanex 923 as extraction reagent. *Desalination* 259 (2010) 179–186.
43. Uğuzdoğan, E.; Denkbaş, E.B.; Kabasakal, O.S. The use of polyethyleneglycolmethacrylate-co-vinylimidazole (PEGMA-co-VI) microspheres for the removal of nickel(II) and chromium(VI) ions. *Journal of Hazardous Materials* 177 (2010) 119–125.
44. Ochoa, N.A.; Illanes, C.; Marchese, J.; Basualto, C.; Valenzuela, F. Preparation and characterization of polymeric microspheres for Cr(VI) extraction. *Separation and Purification Technology* 52 (2006) 39–45.
45. Li, C.; Zhang, Y.; Peng, J.; Wu, H.; Li, J.; Zhai, M. Adsorption of Cr(VI) using cellulose microsphere-based adsorbent prepared by radiation-induced grafting. *Radiation Physics and Chemistry* 81 (2012) 967–970.
46. Zhang, Y.; Xu, L.; Zhao, L.; Peng, J.; Li, C.; Li, J.; Zhai, M. Radiation synthesis and Cr(VI) removal of cellulose microsphere adsorbent. *Carbohydrate Polymers* 88 (2012) 931–938.
47. Zhou, J.; Wang, Y.; Wang, J.; Qiao, W.; Long, D.; Ling, L. Effective removal of hexavalent chromium from aqueous solutions by adsorption on mesoporous carbon microspheres. *Journal of Colloid and Interface Science* 462 (2016) 200–207.
48. Li, P.; Yang, L.; He, X.; Wang, J.; Kong, P.; Xing, H.; Liu, H. Synthesis of PGMA microspheres with amino groups for high-capacity adsorption of Cr(VI) by cerium initiated graft polymerization. *Chinese Journal Chemical Engineering* 20 (2012) 95–104.
49. Wang, Q.; Guan, Y.; Liu, X.; Yang, M.; Ren, X. Micron-sized magnetic polymer microspheres for adsorption and separation of Cr(VI) from aqueous solution. *Chinese Journal of Chemical Engineering* 20 (2012) 105–110.
50. Hua, C.; Zhang, R.; Bai, F.; Lu, P.; Liang, X. Removal of chromium (VI) from aqueous solutions using quaternized chitosan microspheres. *Chinese Journal of Chemical Engineering* 25 (2017) 153- 158
51. Wagdare, N.A.; Baggerman, J.; Marcelis, A.T.M.; Boom, R.M.; Van Rijn, C.J.M. Polymer microspheres with structured surfaces. *Chemical Engineering Journal* 175 (2011) 561–568.
52. Venkateswaran, P.; Palanivelu, K. Solvent extraction of hexavalent chromium with tetrabutyl ammonium bromide from aqueous solution. *Separation and Purification Technology* 40 (2004) 279–284.
53. Hu, B.; Luo, H. Adsorption of hexavalent chromium onto montmorillonite modified with hydroxyaluminum and cetyltrimethylammonium bromide. *Applied Surface Science* 257 (2010) 769–775.
54. Figaro, S.; Avril, J.P.; Brouers, F.; Ouensanga, A.; Gaspard, S. Adsorption studies of molasse's wastewaters on activated carbon: Modelling with a new fractal kinetic equation and evaluation of kinetic models. *Journal of Hazardous Materials* 161 (2009) 649–656.
55. F.C. Wu, R.L. Tseng, R.S. Juang, Initial behavior of intraparticle diffusion model used in the description of adsorption kinetics. *Chemical Engineering Journal* 153 (2009) 1–8.
56. Nghiem, L.D.; Mornane, P.; Potter, I.D.; Perera, J.M.; Cattrall, R.W.; Kolev, S.D. Extraction and transport of metal ions and small organic compounds using polymer inclusion membranes (PIMs). *Journal of Membrane Science* 281 (2006) 7–41.
57. Bilgili, M.S. Adsorption of 4-chlorophenol from aqueous solutions by xad-4 resin: Isotherm, kinetic,

- and thermodynamic analysis. *Journal of Hazardous Materials* 137 (2006) 157–164.
58. Wu, F.C.; Tseng, R.L.; Juang, R.S. Comparisons of porous and adsorption properties of carbons activated by steam and KOH. *Journal of Colloid and Interface Science* 283 (2005) 49–56.
59. Liu, Q.; Zheng, T.; Wang, P.; Jiang, J.; Li, N. Adsorption isotherm, kinetic and mechanism studies of some substituted phenols on activated carbon fibers. *Chemical Engineering Journal* 157 (2010) 348–356.
60. Foo, K.Y.; Hameed, B.H. Insights into the modeling of adsorption isotherm systems. *Chemical Engineering Journal* 156 (2010) 2–10.
61. Milmile, S.N.; Pande, J. V.; Karmakar, S.; Bansiwala, A.; Chakrabarti, T.; Biniwale, R.B. Equilibrium isotherm and kinetic modeling of the adsorption of nitrates by anion exchange Indion NSSR resin. *Desalination* 276 (2011) 38–44.
62. Mahramanlioglu, M.; Kizilcikli, I.; Bicer, I.O. Adsorption of fluoride from aqueous solution by acid treated spent bleaching earth. *Journal of Fluorine Chemistry* 115 (2002) 41–47.

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Ouazine L., Bey S., Benamor M., Preparation and characterization of PVC/Aliqat-336 microspheres for chromium (VI) removal, ***Algerian J. Env. Sc. Technology*, 8:1 (2022) 2271-2284**